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Technical Memorandum DRAFT

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Technical Memorandum 1

Subject: Evaluation of Ammonia Toxicity during Elk River Wastewater Effluent Mixing in Humboldt Bay

Date: November 25, 2019

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Limitations:

This is a draft memorandum and is not intended to be a final representation of the work done or recommendations made by Brown and Caldwell. It should not be relied upon; consult the final report.

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Executive Summary

The North Coast Regional Water Quality Control Board (Water Board) has decided that the City of Eureka (City) Elk River Wastewater Treatment Plant (WWTP) effluent discharge to Humboldt Bay should be considered as a bay rather than an ocean discharge. As such Water Board staff has proposed a stringent limit for effluent ammonia owing to concerns about un-ionized ammonia toxicity. The reader should note that the El River WWTP discharge differs significantly from any other Humboldt Bay discharge (e.g., compared to the very-shallow-water Arcata discharge) or indeed, any discharge elsewhere in Region 1.

To check for toxicity from the City's discharge, Brown and Caldwell (BC) modeled mixing at flow rates of 6 million gallon per day (mgd) and 30 mgd, and then carried out chemical modeling for ammonia toxicity for the 6-mgd discharge rate using effluent data from a fall 2019 discharge day and the worst-case receiving water conditions (no current across the diffuser plus late summer/early fall receiving water conditions) for potential toxicity. This technical memorandum (TM) presents results of analyses that show no un-ionized ammonia toxicity at any time during the discharge. Even at much higher effluent total ammonia concentrations, un-ionized ammonia toxicity would not occur.

Section 1: Introduction and Background

This TM presents analyses to evaluate possible ammonia toxicity for effluent discharged from the City WWTP as it mixes in Humboldt Bay. The TM presents background information about the discharge, estimates for its mixing after discharge and chemical modeling to check for ammonia toxicity.

1.1 Background Information

Since the early 1980s, the City has discharged secondary effluent from the Elk River WWTP through a submerged multiport diffuser in Humboldt Bay (Bay). When the City planned the facilities, the best available information indicated that effluent discharged on the outgoing tide would leave the Bay. Hence, for its discharge permit, the Water Board allowed discharge categorization as an ocean discharge. As an ocean discharger, the WWTP received credit for the initial dilution (30 parts seawater per part wastewater), which resulted in a six-month median ammonia limit of 18.6 milligrams per liter (mg/L) (Water Board, 2009, order R1-2009-0033).

More recent discharge modeling by Humboldt State University (Brown and Caldwell, 2014) has shown that even with a carefully timed discharge strategy, not all effluent exits the Bay. Hence, the Water Board has determined that the City's discharge is a bay discharge governed under the California Toxics Rule and Water Quality Control Plan for the Enclosed Bays and Estuaries of California Plan (Water Quality Control Plan for Enclosed Bays and Estuaries of California). The most recent National Pollutant Discharge Elimination System (NPDES) permit (R1-2016-0001) established limits for ammonia, without the benefit of a mixing zone or dilution credit, because the Basin Plan authorizes neither. Based on the Ambient Water Quality Criteria for Ammonia (Saltwater)-1989 (United States Environmental Protection Agency (USEPA), April 1989), the unionized ammonia criteria concentration is 0.035 mg/L (four-day average concentration); Water Board staff used tables from this document to calculate criteria concentrations for total ammonia based on receiving water temperature and pH. Recent communications with Water Board staff have indicated that the next permit will contain revised ammonia criteria concentrations based on the following assumptions and a Water Board staff toxicity evaluation:

- Receiving water pH of 8.2 (maximum measured receiving water pH was 8.1)
- Receiving water temperature of 15°C (maximum measured temperature was 14 degrees Celsius (°C)



The new total ammonia limits would be 1.68 mg/L as nitrogen (N) maximum day effluent limit and 0.9 mg/L as N average monthly effluent limit.

1.2 Scope of Work

Through its contract with GHD Engineering, the City directed BC to evaluate ammonia toxicity for effluent as it mixes in the Bay. Work included dilution calculations, chemical modeling, comparison with USEPA criteria, and reporting of findings and conclusions.

Section 2: Bay Outfall Dilution Modeling

BC evaluated predicted dilution for the existing effluent discharge, with proposed discharge port improvements, to Humboldt Bay using the outfall dilution model UM3. UM3 is a model in the most recent release of the USEPA-supported Visual Plumes modeling package (https://www.epa.gov/ceam/visual-plumes). Model analyses focused on initial, jet momentum-based dilution achieved within the first few seconds after discharge and assumed worst case ambient conditions. BC also carried modeling results through initial dilution completion. The predicted dilutions provide input to water quality toxics analyses for ammonia presented in Section 3.

2.1 Model Description and General Plume Mixing Concepts

Visual Plumes is a Windows-based graphical user interface to a suite of numerical plume models. UM3 is a three-dimensional initial dilution plume model applicable to submerged single and multi-port diffusers, capable of modeling both positively and negatively buoyant plumes. BC selected Visual Plumes for mixing modeling since it is well proven and widely used in California and is appropriate for the type of discharge and receiving water conditions.

Typically, the wastewater industry and regulators describe mixing of effluent discharged from an outfall to receiving waters in two distinct phases: 1) rapid initial dilution in the near-field, and 2) slower subsequent dilution in the far-field. Rapid initial dilution in the near-field has two distinct physical components. The first component is turbulent jet mixing and entrainment resulting from discharge momentum while exiting the diffuser ports. The second component is turbulent mixing and entrainment resulting from the plume rising in the water column due to the effluent's buoyancy. When the jet momentum and buoyancy-mixing forces dissipate, the slower process of subsequent dilution continues in the far-field. Mixing and dispersion in the far-field occurs along the boundaries of the plume, primarily in the horizontal plane laterally and longitudinally as the plume is carried by ambient currents. For the City's discharge, subsequent dilution is not relevant for the toxicity questions and BC did not considered it further for this TM.

The UM3 model reports the initial dilution value as either "center-line" or "flux-average". The flux-average dilution (the average dilution across the plume width) is always greater than the center-line dilution (the minimum dilution in the middle of the plume). Typically, one would use a center-line dilution for analysis of acute water quality concerns, while the flux-average dilution is more appropriate for chronic conditions. For this TM, we focus on the more conservative (i.e., higher) predicted centerline values.

2.2 Outfall Dilution Model Input Parameters

Input parameters to the UM3 model include the outfall discharge physical configuration, and effluent and receiving water characteristics including flow, temperature, and salinity.



2.2.1 Outfall Dimensions

Elk River WWTP effluent discharges to Humboldt Bay through an outfall structure consisting of a 48-inch-diameter pipe, 4,100 feet in length, terminating in a multi-port diffuser. The existing diffuser has 90, 3-inch-diameter ports, with a port spacing of 4 feet. The originally installed pipe had flaps over the ports. Recent diver inspection has determined that the flaps largely have failed. Proposed discharge port improvements include installation of elastomeric Tideflex® check valves (manufactured by Red Valve, Inc., Pittsburgh, PA), that would improve hydraulic and dilution performance over the range of anticipated effluent flows while preventing detrimental sediment intrusion into the diffuser during lower flows. Attachment A contains existing outfall drawings and proposed discharge port improvements. Based on manufacturer data, the proposed valves have an effective diameter of 1.9 inches at an effluent flow of 6 mgd and 2.9 inches at an effluent flow of 30 mgd as discussed further below.

The effluent ports discharge horizontally relative to the Bay bottom, aligned parallel to the prevailing current regime (vertical and horizontal angles = 0 degrees). Port depth varies within the tidal cycle. The model analyses herein assume a depth of 25 feet based on average tidal conditions, but initial jet mixing, likely the most critical condition for ammonia toxicity, is independent of depth.

2.2.2 Effluent Characteristics

BC modeled effluent flow rates of 6 and 30 mgd. We selected 6 mgd to represent flow at a worst-case dry weather condition as would occur in late summer/early fall before the wet season commences. Bay waters are warmest then. Based on City data, we set effluent temperature at 20°C, with a salinity of 0 practical salinity units (psu).

2.2.3 Receiving Water Characteristics

Receiving water characteristics include ambient salinity and temperature profiles for the water column, and current speed and direction with respect to the outfall discharge.

- The receiving water would have uniform salinity and temperature characteristics throughout the water column (unstratified conditions) (Brown and Caldwell 2014 and Brown and Caldwell 1981). Data reported at the Central and Northern California Ocean Observing System station at Humboldt Bay indicate typical average temperature and salinity conditions within the Bay are 15 °C and 34 psu, respectively. The temperature assumption is consistent with the receiving water temperature assumed by Water Board staff to calculate revised ammonia criteria concentrations for the next permit.
- Per the typical assumption for dilution analyses in California, we set the current speed conservatively at zero, consistent with discharge initiated at slack water. Discharge flow is parallel to the prevailing current.

2.3 Outfall Dilution Model Results

We performed UM3 model runs for effluent flow rates of 6 mgd and 30 mgd, with corresponding port diameters as described above. We selected 6 mgd as representative of dry weather discharges and 30 mgd as representative of peak wet weather flow. Tables 2-1 and 2-2 present results for each modeled flow rate, including model predicted dilution, plume depth, and travel time at modeled steps (distance) from the diffuser ports. Tabular results also include distances at which the individual port discharge plumes merge and rise to the water surface due to buoyancy effects. Predicted dilution is a dilution factor calculated as the effluent plume volume plus the entrained ambient volume, divided by the effluent plume volume. A dilution factor of 1 represents 100 percent effluent.

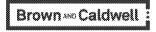
Figures 2-1 and 2-2 present tabular modeling results graphically. Attachment B provides detailed model input and output data.



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	Table 2-1. Model Re	sults Summary, Effluen	i Flow = 6 mgd	
Distance from Diffuser (feet)	Predicted Dilution	Plume Depth (feet)	Travel Time (seconds)	Plume Notes
0.3	1.8	25	0.1	
0.9	3.3	25	0.4	
1.8	5.6	25	1.2	
2.6	7.6	25	2.2	
3.2	9.5	24	3.3	
3.9	12	24	4.6	
4.5	14	23	6.2	
5.2	17	23	8.3	
6.0	23	21	12	
7.1	34	19	17	
8.2	53	15	27	Plume merges
8.5	59	14	30	
10.1	107	2.7	55	
10.3	118	0.2	61	Plume surfaces

	Table 2-2. Model Re	sults Summary, Effluent	Flow = 30 mgd	
Distance from Diffuser (feet)	Predicted Dilution	Plume Depth (feet)	Travel Time (seconds)	Plume Notes
0.5	1.9	25	0.1	
1.4	3.3	25	0.3	
3.0	6.0	25	1.0	
5.5	10	25	2.8	
7.5	14	24	4.9	
9.2	17	23	7.3	
9.6	18	23	8.0	Plume merges
10.8	20	22	10	
12.6	23	20	14	
14.7	28	18	19	
17.3	35	14	26	
20.9	48	4.7	39	
21.4	51	2.9	42	
22.1	54	0.8	45	Plume surfaces



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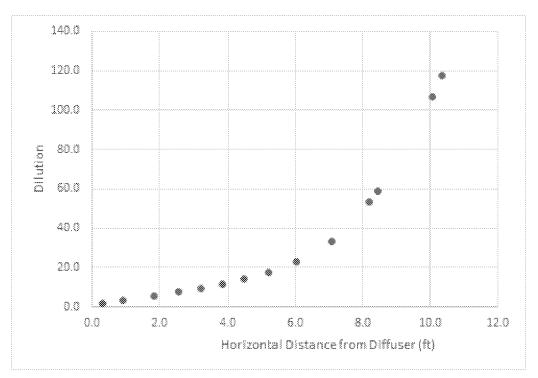


Figure 2-1. Dilution versus horizontal distance from diffuser, effluent flow = 6 mgd

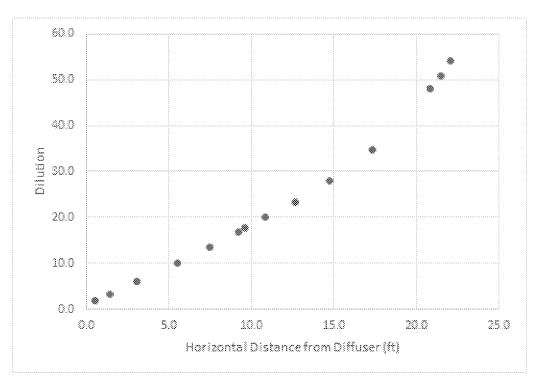
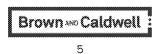


Figure 2-2. Dilution versus horizontal distance from diffuser, effluent flow = 30 mgd



2.4 Outfall Dilution Modeling Conclusions

Model results provided in Tables 2-1 and 2-2 show that for both the 6 mgd and 30 mgd cases, rapid effluent dilution of 10:1 occurs three to four seconds following discharge and within 25 port diameters (less than six feet for both cases) of the diffuser. Negligible buoyancy-based plume rise occurs within this short distance; this initial dilution is due almost entirely to jet momentum-based entrainment of the receiving water. The concentrated effluent plume, at dilutions less than 10:1, is limited to a very small bottom area near the diffuser. This area is not continuous along the diffuser axis, as the plumes from individual ports have not yet merged, i.e., it is confined to a small area around each port.

The discharge achieves much higher slack-water dilutions (in excess of 100 for an effluent flow of 6 mgd and in excess of 50 for an effluent flow of 30 mgd), as the buoyancy-based effluent mixing continues and the plume entrains additional receiving water at distances further from the diffuser. The model predicts that the effluent plume surfaces between 10 and 22 feet from the diffuser, for the 6 mgd and 30 mgd cases, respectively.

Section 3: Chemical Modeling

BC used the widely applied OLI Studio electrolyte thermodynamic software to calculate water quality in blends of wastewater effluent with the seawater at various blending ratios. BC carried out chemical modeling only for the low-flow condition, since at high flows the effluent ammonia concentration is lower because of dilute influent.

3.1 Chemical Model Inputs

Table 3-1 lists water quality parameters used as inputs to the model. To model the changes in pH, concentrations of all of the cations and anions were needed. Since these data are not routinely collected, BC extracted wastewater effluent concentration data from an Elk River WWTP wastewater effluent sample collected October 3, 2019 and analyzed by North Coast Laboratory. Wastewater effluent temperature and pH were assumed to be the yearly average values.

Receiving water temperature and pH were assumed to be consistent with the values assumed by Water Board staff to calculate revised ammonia criteria concentrations for the next permit. Cation and anion data were not available for the receiving water, so standard seawater composition was obtained from a literature review (see Table 3-1).

Brown № Caldwell

Table 3-1. Chemic	al Model Inputs for t	Vastewater Effluent and Sea	Water Composition
Parameters	Units	Wastewater Effluent	Seawater ^{a,b}
Alkalinity	mg/L as CaCO ₃	93.0	103.9
Conductivity	μS/cm	940	N/A
рН	s.u.	6.5	8.20
Silica	mg/L	13.0	N/A
Temperature	°C	17.7	15.0∘
Cations			
Ammonia	mg/L as N	9.3	N/A
Calcium	mg/L	33.0	424
Magnesium	mg/L	15.0	1,331
Potassium	mg/L	24.0	411
Sodium	mg/L	66.0	11,083
Anions			
Bicarbonate	mg/L as CaCO₃	93.0	N/A
Chloride	mg/L	150	19,934
Fluoride	mg/L	0.22	N/A
Nitrate	mg/L as NO₃	42.1	N/A
Ortho-phosphate	mg/L as HPO4b	16.7	N/A
Sulfate	mg/L	38.0	2,793

- a. Alkalinity value for seawater was calculated in the model. CaCO₃ is calcium carbonate
- b. Receiving water cation and anion composition was not available, so seawater composition obtained from

http://www.soest.hawaii.edu/oceanography/courses/OCN623/Spring2012/Salinity2012web.pdf

c. Receiving water temperature and pH were consistent with values assumed by the Water Board

 μ S/cm = microSiemens per centimeter

CaCO3 = calcium carbonate

N/A = not applicable $NO^3 = nitrate$

s.u. = standard units

3.2 Chemical Model Results

Table 3-2 presents the model output and compares the predicted ammonia concentrations (both unionized and total) to the USEPA 1989 4-day average criteria. Based on the modeling results, un-ionized ammonia is less than the 4-day average criterion (0.035 mg/L un-ionized ammonia) at a range of dilution ratios (see Figure 3-1). Figure 3-2 shows the results for total ammonia and pH. The criteria shown, which vary with pH, temperature, and salinity, are not exceeded at any dilution ratios. The USEPA 1989 one-hour average criteria are much higher than the 4-day average criteria, and so would not be exceeded either.

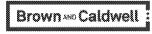


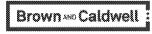
				Table	3-2. Chemi	sal Model Outpul				
				Uı	n-ionized Am	monia		Total Amn	nonia	
Dilution Ratio	рН	Temperature °C	Salinity g/kg	In mixture (mg/L)	Criterion (mg/L)	Criterion Exceeded?	In mixture (mg/L as N)	In mixture (mg/L)	Criterion (mg/L)ª	Criterion Exceeded ?
1	6.5	18	0.27	0.011	0.035	NO	9.32	12.0	NA	NA
1.83	6.5	17	14.9	0.005	0.035	NO	5.45	7.02	9.4	NO
2	6.6	16	17.8	0.0038	0.035	NO	4.66	6.00	9.7	NO
3	6.7	16	23.6	0.0034	0.035	NO	3.11	4.00	9.7	NO
3.3	6.8	16	24.5	0.0035	0.035	NO	2.87	3.70	9.7	NO
4	6.9	16	26.5	0.0035	0.035	NO	2.33	3.00	9.7	NO
5	7.0	16	28.2	0.0037	0.035	NO	1.86	2.40	6.2	NO
5.6	7.1	16	28.9	0.0039	0.035	NO	1.68	2.16	6.2	NO
6	7.1	16	29.3	0.0041	0.035	NO	1.55	2.00	6.2	NO
7	7.3	15	30.1	0.0045	0.035	NO	1.33	1.71	4.1	NO
7.6	7.3	15	30.5	0.0048	0.035	NO	1.23	1.59	4.1	NO
8	7.4	15	30.7	0.0049	0.035	NO	1.16	1.50	4.1	NO
9	7.5	15	31.2	0.0054	0.035	NO	1.04	1.33	3.1	NO
9.5	7.5	15	31.4	0.0056	0.035	NO	0.98	1.27	3.1	NO
10	7.5	15	31.6	0.0059	0.035	NO	0.93	1.20	3.1	NO
11	7.6	15	31.9	0.0062	0.035	NO	0.85	1.09	3.1	NO
11.5	7.6	15	32.0	0.0064	0.035	NO	0.81	1.05	1.7	NO
12	7.7	15	32.1	0.0065	0.035	NO	0.78	1.00	1.7	NO
13	7.7	15	32.4	0.0067	0.035	NO	0.72	0.92	1.7	NO
14	7.8	15	32.5	0.0068	0.035	NO	0.67	0.86	1.7	NO
15	7.8	15	32.7	0.0069	0.035	NO	0.62	0.80	1.7	NO
17	7.8	15	33.0	0.0069	0.035	NO	0.55	0.71	1.1	NO
17.4	7.9	15	33.0	0.0069	0.035	NO	0.54	0.69	1.1	NO
19	7.9	15	33.2	0.0068	0.035	NO	0.49	0.63	1.1	NO
21	7.9	15	33.4	0.0067	0.035	NO	0.44	0.57	1.1	NO
22.8	8.0	15	33.5	0.0065	0.035	NO	0.41	0.53	1.1	NO
23	8.0	15	33.5	0.0065	0.035	NO	0.41	0.52	1.1	NO
25	8.0	15	33.6	0.0063	0.035	NO	0.37	0.48	1.1	NO
27	8.0	15	33.7	0.0061	0.035	NO	0.35	0.44	1.1	NO
29	8.0	15	33.8	0.0059	0.035	NO	0.32	0.41	0.69	NO
31	8.0	15	33.9	0.0054	0.035	NO	0.30	0.39	0.69	NO
33	8.0	15	33.9	0.0054	0.035	NO	0.28	0.36	0.69	NO
33.5	8.0	15	34.0	0.0054	0.035	NO	0.28	0.36	0.69	NO



				Table	3-2. Chemi	eal Wodel Output				
				Uı	n-ionized Am	monia		Total Amn	nonia	
Dilution Ratio	рН	Temperature °C	Salinity g/kg	In mixture (mg/L)	Criterion (mg/L)	Criterion Exceeded?	In mixture (mg/L as N)	In mixture (mg/L)	Criterion (mg/L)ª	Criterion Exceeded ?
35	8.0	15	34.0	0.0053	0.035	NO	0.27	0.34	0.69	NO
37	8.1	15	34.0	0.0051	0.035	NO	0.25	0.32	0.69	NO
39	8.1	15	34.1	0.0049	0.035	NO	0.24	0.31	0.69	NO
41	8.1	15	34.1	0.0047	0.035	NO	0.23	0.29	0.69	NO
46	8.1	15	34.2	0.0044	0.035	NO	0.20	0.26	0.69	NO
51	8.1	15	34.3	0.0040	0.035	NO	0.18	0.24	0.69	NO
53.3	8.1	15	34.3	0.0039	0.035	NO	0.18	0.23	0.69	NO
56	8.1	15	34.4	0.0038	0.035	NO	0.17	0.21	0.69	NO
58.9	8.1	15	34.4	0.0036	0.035	NO	0.16	0.21	0.69	NO
69	8.1	15	34.5	0.0032	0.035	NO	0.14	0.17	0.69	NO
82	8.1	15	34.5	0.0027	0.035	NO	0.11	0.15	0.69	NO
95	8.1	15	34.6	0.0024	0.035	NO	0.10	0.13	0.69	NO
106.6	8.2	15	34.6	0.0022	0.035	NO	0.09	0.11	0.69	NO
108	8.2	15	34.6	0.0022	0.035	NO	0.09	0.11	0.69	NO
117.7	8.2	15	34.7	0.0020	0.035	NO	0.08	0.10	0.69	NO
120	8.2	15	34.7	0.0020	0.035	NO	0.08	0.10	0.69	NO

a. Criterion for total ammonia are from tables in Ambient Water Quality Criteria for Ammonia (Saltwater)-1989 (EPA-440/5-88-004, April 1989) for the modeled pH, temperature, and salinity. Temperature was rounded up to the nearest 5 °C, and pH was rounded up to the nearest 0.2.

g/kg = gram per kilogram



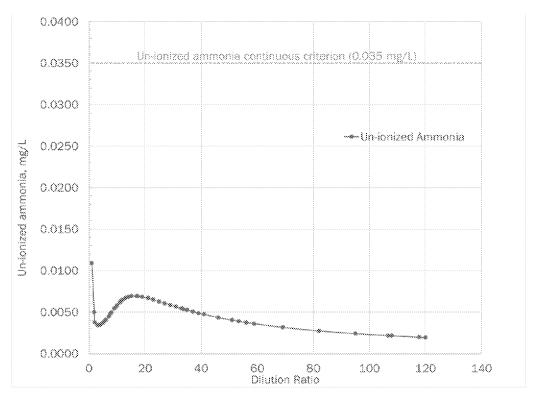


Figure 3.1: -Unionized ammonia at various dilution ratios

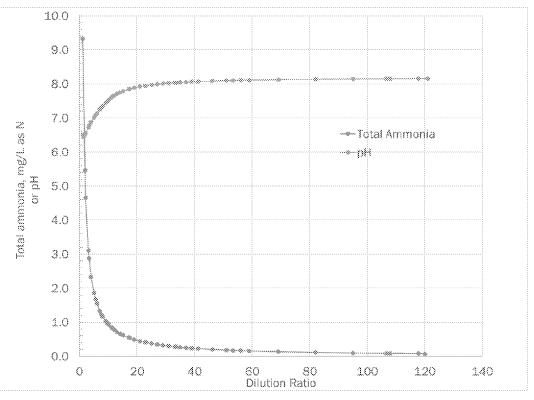
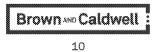


Figure 3-2. Total ammonia concentration (as N) and pH at various dilution ratios



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Effluent Mixing Draft TM 2019-11-26 docx

3.3 Chemical Modeling Discussion

The results presented above show that at no time during the discharge described in this TM would the effluent plume be toxic for un-ionized ammonia. However, during the period September 2016 through March 2019, the highest discharge effluent total ammonia concentration was 18 mg/L. As Figure 3-1 shows, even at a much higher effluent ammonia concentrations than modeled, the effluent would continue to be nontoxic for unionized ammonia.

References

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North Coast Regional Water Quality Control Board (Water Board), Waste Discharge Requirements for the City of Eureka, Elk River Wastewater Treatment Plant, Humboldt County, Order No. R1-2009-0033, NPDES No. CA0024449, June 2009.



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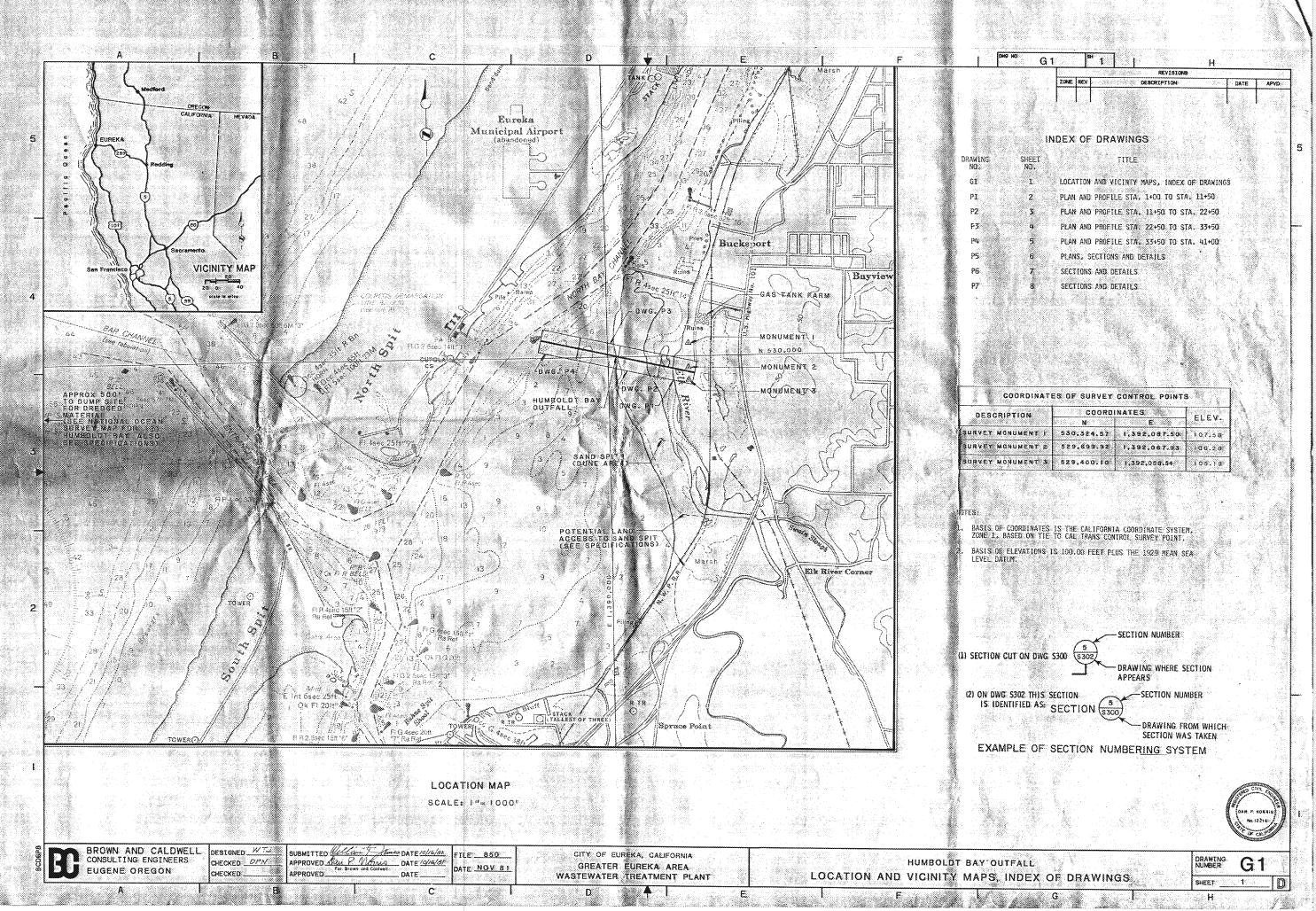
Attachment A: Existing Outfall Drawings and Proposed Discharge Port Improvements

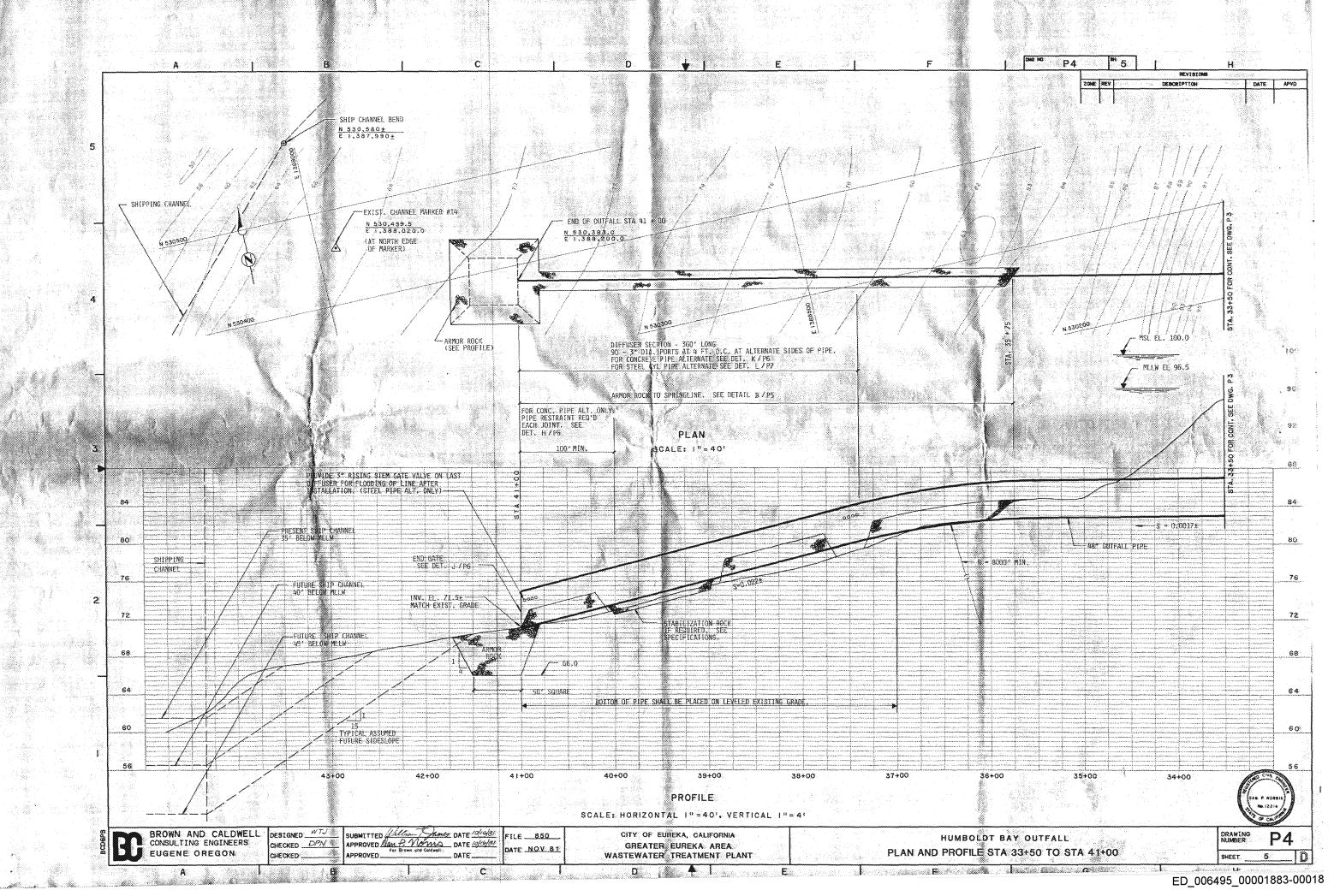


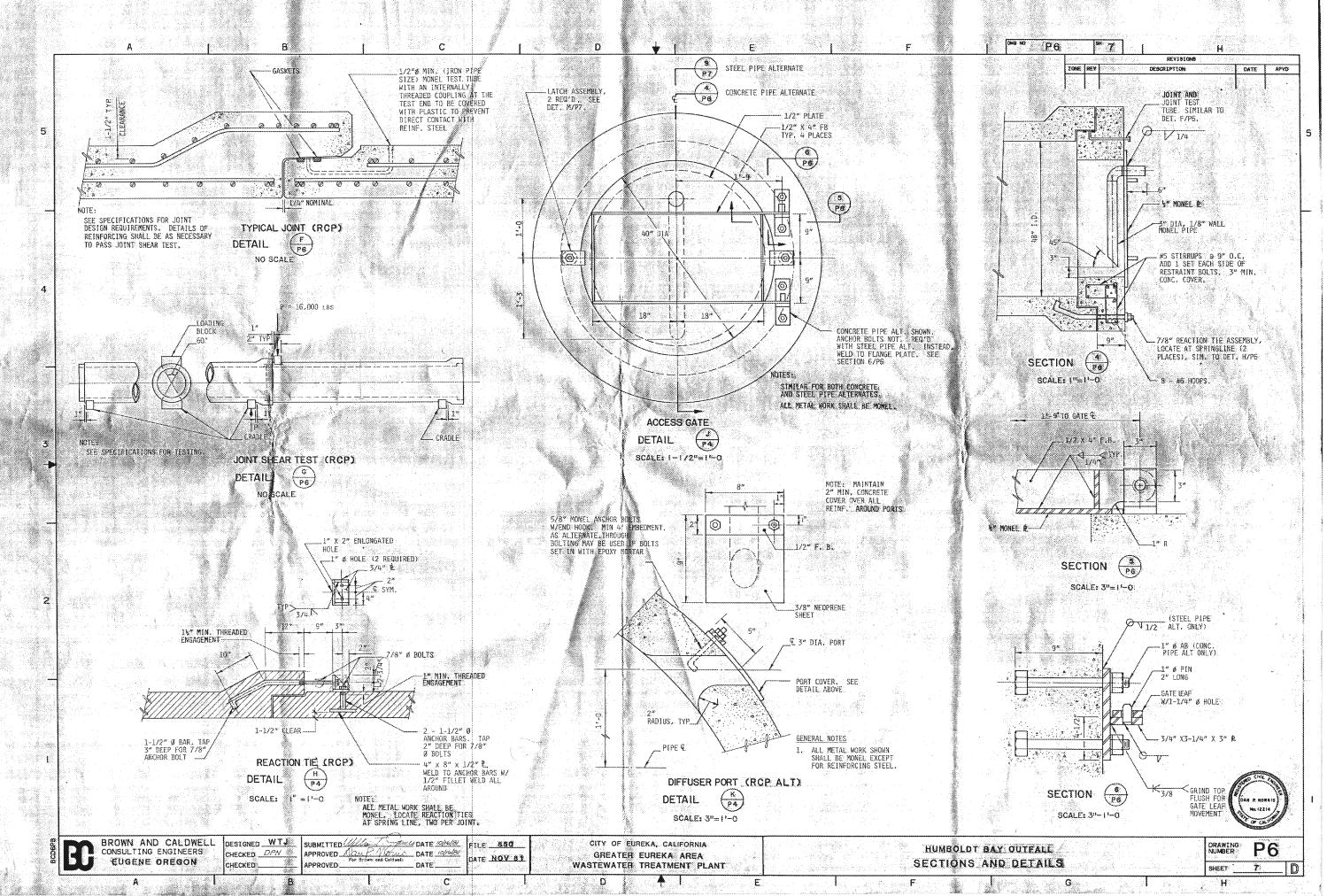
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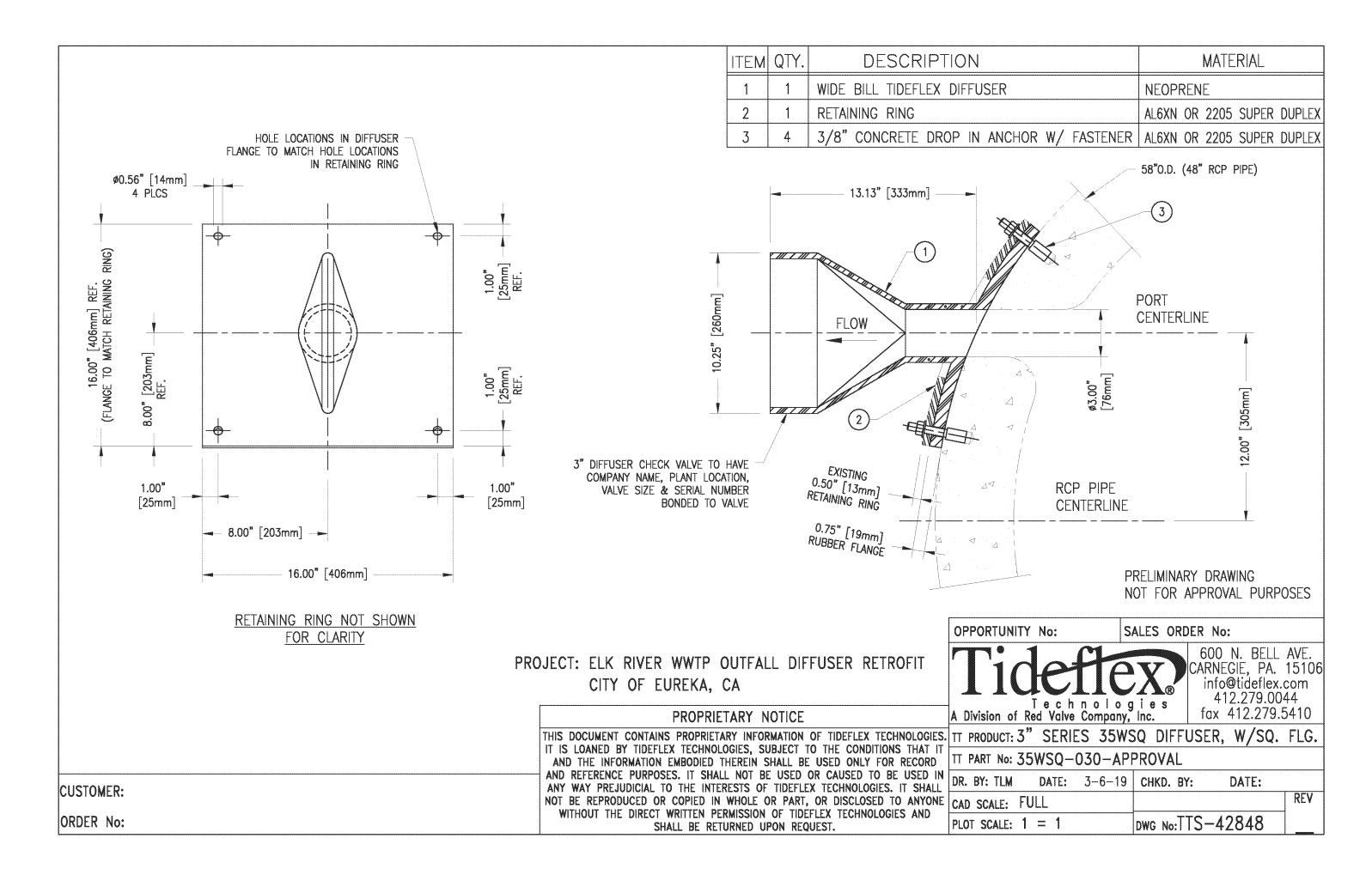


Α









Attachment B: Detailed UM3 Model Input and Output Data



Technical Memorandum 1	Evaluation of Ammonia Toxicity during Elk River Wastewater Effluent Mixing in Humboldt Bay
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Eureka Model Run: 6mgd

Ambient	Table:												
Depth	Amb-cur	Amb-dir	Amb-sal	. Amb-ter	n Amb-	pol	Decay :	Far-spd	Far-dir	Disprsn	Densi	ty	
m	m/s	deg	psu	С	kg/	'kg	s-1	m/s	deg	m0.67/s2	2 sigm	a-T	
0.0	0.0	0.0	34.00	15.00	_ C	0.0	0.0	-		0.0003	25.22	390	
1.219	0.0	0.0	34.00	15.00	C	0.0	0.0	_	_	0.0003	25.22	390	
2.438	0.0	0.0	34.00	15.00	C	0.0	0.0	-	_	0.0003	25.22	390	
3.658	0.0	0.0	34.00	15.00	C	0.0	0.0	_	_	0.0003	25.22	390	
4.877	0.0	0.0	34.00	15.00	C	0.0	0.0	_	_	0.0003	25.22	390	
6.096	0.0	0.0	34.00	15.00	C	0.0	0.0	_		0.0003	25.22	390	
7.315	0.0	0.0	34.00	15.00	C	0.0	0.0	_	_	0.0003	25.22		
	er table:												
	Ver angl			SourceY	Ports	Spacing		Isoplth					Polutnt
(in)	(deg)	(deg)	(m)	(m)	()	(ft)	(ft)	(concent)		(MGD)	(psu)	(C)	(용)
1.900	0.0	0.0	0.0	0.0	90	4.00	30.00	0.0	25.0	6.00	0.0	20.00	100.00
Simulat													
SIMUIA		7 mla augus	P-dia	Polutnt	Dilutn			n Time	Iso dia	-			
C +	Depth	Amb-cur				x-posn				i			
Step	(ft) 25.00	(cm/s) 0.001	(in)	(%) 100.0	()	(ft) 0.0	(ft) 0.0	(s) 0.0	(m) 0.0	_			
0		0.001	1.900 1.937		1.000						1-2		
1	25.00			96.18	1.040	0.0151	0.0			; bottom	nit;		
30	25.00	0.0	3.456 6.234	54.77	1.826	0.324	0.0		0.08777				
60 90	24.98	0.0	10.65	30.43 17.76	3.286	0.901	0.0		0.1583 0.2705				
	24.86	0.0			5.631	1.819							
120	24.66	0.0	14.21	13.09	7.639	2.577	0.0		0.3609				
150	24.37	0.0	17.14	10.52	9.509	3.230	0.0		0.4354				
180	23.97	0.0	19.83	8.677	11.53	3.854	0.0		0.5037				
210	23.41	0.0	22.57	7.153	13.98	4.496	0.0		0.5733				
240	22.57	0.0	25.72	5.756	17.37	5.202	0.0		0.6533				
270	21.23	0.0	30.00	4.386	22.80	6.031	0.0		0.7620				
300	18.77	0.0	37.16	2.990	33.45	7.082	0.0		0.9439				
325	14.77	0.0	48.48	1.875	53.33	8.203	0.0			; merging	J;		
330	13.62	0.0	51.33	1.698	58.88	8.452	0.0		1.3037				
360	2.728	0.0	73.95	0.938	106.6	10.06	0.0		1.8784		_		_
361	2.234	0.0	74.96	0.919	108.8	10.12	0.0	56.27	1.9041	; matched	d energy	radial	.vel =

	Model Ru t Table:	ın: 30mgd											
Depth	Amb-cur	Amb-di	r Amb-	-sal Amb	o-tem 1	Amb-pol	Decay	Far-spd	Far-dir	Disp	rsn De	ensity	
m	m/s	deg	ps		С	kg/kg	s-1	m/s	deq	m0.67		igma-T	
0.0	0.0	0.0	34.		5.00	0.0	0.0	-	_	0.000		22390	
1.219	0.0	0.0	34.	00 1	5.00	0.0	0.0	_	_	0.000	3 25.	22390	
2.438	0.0	0.0	34.	00 1	5.00	0.0	0.0	_	_	0.000	3 25.	22390	
3.658	0.0	0.0	34.	00 1	5.00	0.0	0.0	-	_	0.000	3 25.	22390	
4.877	0.0	0.0	34.	00 1	5.00	0.0	0.0	-	-	0.000	3 25.	22390	
6.096	0.0	0.0	34.	00 1	5.00	0.0	0.0	_	_	0.000	3 25.	22390	
7.315	0.0	0.0	34.	00 1	5.00	0.0	0.0	-		0.000	3 25.	22390	
D: 55	er table:												
	er tabre: Ver angl		SourceV	Source:	Y Ports	Chadina	M7 dia	Isoplth	D donth	m+1 f10	Def anl	Temp	Polutnt
r-ura (in)	(deg)	(deg)	(m)	m)	()	(ft)	(ft)	(concent)	(ft)	(MGD)	(psu)	(C)	(%)
2.900	0.0	0.0	0.0	0.0	90	4.00	30.00	0.0	25.0	30.00	0.0	20.00	100.00
2.500	0.0	0.0	0.0	0.0	20	1.00	30.00	0.0	20.0	30.00	0.0	20.00	100.00
	-	Amb-cur		Polutnt	Dilutn	x-posn	y-posr		Iso di				
Step	(ft)	(cm/s)	(in)	(용)	()	(ft)	(ft)		(n				
0	25.00	0.001	2.900	100.0	1.000	0.0							
1	25.00	0.0	2.980	94.72	1.056	0.0319				69; bott	om hit;		
30	25.00	0.0	5.358	53.93	1.854	0.512				361;			
60	24.99	0.0	9.674	29.96	3.338	1.406							
90	24.91	0.0	17.47	16.60	6.025	3.028							
120	24.57	0.0	29.25	9.866	10.14	5.482							
150	23.99	0.0	38.47	7.362	13.58	7.454							
180	23.19	0.0	46.16	5.934	16.85	9.175							
188	22.93	0.0	48.09	5.632	17.76	9.618				215; merg	ing;		
210	22.08	0.0	52.78	4.986	20.05	10.84							
240	20.45	0.0	58.78	4.266	23.44	12.63							
270	17.91	0.0	65.53	3.584	27.90	14.70							
300	13.53	0.0	75.12	2.869	34.85	17.29							
330	4.729	0.0	93.21	2.077	48.14	20.85							
334	2.876	0.0	97.02	1.965	50.90	21.44	0.	0 42.07	2.46	343; matc	hed ener	gy radi	.al vel =